

# Effective Biosorptive Elimination of Fluoride and Sulfide Ions from Water Using Chemical Modified Rice Bran in Economical Way

# AMARA DAR<sup>1</sup>, RABIA REHMAN<sup>1</sup>\*, PARSA DAR<sup>2</sup>, AYESHA KAUSAR<sup>3</sup>, SABA NOREEN<sup>4</sup>, LIVIU MITU<sup>5</sup>\*

<sup>1</sup>Institute of Chemistry, University of the Punjab, Lahore-54590, Pakistan

<sup>2</sup>University of Macau, Taipa, Macau, China

<sup>3</sup>Shanghai Jiao Tong University, Shanghai, China

<sup>4</sup>University of the Chinese Academy of Sciences Beijing, Beijing, China

<sup>5</sup>University of Pitesti, Department of Nature Sciences Pitesti, 1 Targu din Vale Str., 110040, Pitesti, Romania

Abstract. Due to industrial development the sulfide ions are present in excess in the effluents from tanneries, whereas ceramic and glass industries are the major contributors of fluoride ions. The concentration of these ions is beyond permissible limit in water bodies, thereby causing serious health issues. An attempt is made to highlight the effectiveness of chemical modification of rice bran with zinc chloride for the removal of sulfide and fluoride ions from aqueous solutions. Batch mode adsorption study was conducted. Mechanism of adsorption was explored by applying the isothermal models like Langmuir and Freundlich. Maximum adsorption capacity (mg/g) revealed that zinc chloride modification of rice bran increased the sulfide and fluoride removal from 3.79 mg/g to 7.08 mg/g and 1.37 mg/g to 2.06 mg/g, respectively. Sulfide ions were more efficiently eliminated using the selected adsorbent in its modified form.

Keywords: Sulfide, Fluoride, Biosorption, Rice bran, Freundlich and Langmuir isotherms.

# **1.Introduction**

To meet the requirements of increasing population, our industrial sector has been developing with same pace. Industries are producing waste in both liquid and solid forms. The pollution in our aquatic environment is a big concern because the pollutants in form of metal ions are directly or indirectly threatening to environment and human life [1]. As the conventional treatment methods are not cost effective and low-cost treatment methods and technologies are unable to remove wishful amount of pollutants so, industrial waste water released directly into drain or nearby land devoid of any pre-treatment.

Fluoride as an essential part of drinking water is constructive in making our teeth and bones healthy, while intake of fluoride in excess may inflames skeletal fluorosis, which results in bone deformities [2]. Besides of water, which is being used for drinking purpose, fluoride can also enter our body through industrial exposure. The waste water of fertilizer [3], ceramics [4], glass [5], steel production industry [6] and the plants that are incinerate municipal waste [7] are the major source of fluoride pollution in aquatic environment.

Sulfide ( $S^{2-}$ ), a divalent anion of sulfur, is present in ground water because of sulfide ores. Sulfide ions either in form of H<sub>2</sub>S exposure or other water-soluble sulfide salts are very harmful to our body [8]. Nowadays many industries like pulp paper manufacturers, petrochemical and the most important tanneries (to transforms animal hide into leather) are using sulfide species [9]. So, tanneries and other industrial effluents are considered as main anthropogenic sources that hold a bulk amount of sulfide ions.

<sup>\*</sup>email: grinorganic@yahoo.com; ktm7ro@yahoo.com;



There were various technologies and scientific methods that had been practiced removing these toxic species from water [10]. To control the concentration of fluoride in aquatic environment scientists have used reverse osmosis, coagulation, electrocoagulation [11], precipitation [12], nanofiltration [13], electro-dialytic membrane technologies [14, 15]

and adsorption [16] using various materials [17]. On the other hand, sulfide concentration on a large scale can be controlled by the process of aeration and by using chemical oxidants [18] like Cl<sub>2</sub>, KMnO<sub>4</sub>, H<sub>2</sub>O<sub>2</sub>, ClO<sup>-</sup>, O<sub>2</sub>, manganese (IV)-oxide (MnO<sub>2</sub>) [10, 19]. Surrounded by the aforesaid technologies and methods, biosorption is more effective one for controlling the alarming concentration of fluoride and sulfide in aquatic environment, mainly in waste water of industries. For adsorption of various toxic metals and ions, plant wastes are very important candidate from last few decades [20]. As plant waste is inexpensive, freely available, and has good and selective adsorption capacity for various pollutants [21, 22]. The process of adsorption solely depends on ions of choice (adsorbate) [23] which diffuse to the superficial surface of adsorbent, where weak intermolecular forces incorporate to hold adsorbate with adsorbent [24]. By chemical treatment of plant waste or by employing various modifying agents (base solution of sodium hydroxide) [25, 26]; the biosorption capacity can be enhanced to remove sulfide and fluoride species [27]. Rice bran, a hard-outer layer and byproduct of rice (Oryza sativa L.) has been used in present study. Rice is an Asian crop with 90 % production and intake in this expanse. With Pakistan reference, this country is 5th largest exporter and at 11<sup>th</sup> in world rice production so, rice bran is easily available when brown rice travels through the whiteners and polishers. In our present study, we have used the conventional method of adsorption because of its various practiced characteristics like selectivity, suitability, regenerability and compatibility with our lab facilities. Rice bran has been used as an adsorbent (phyto-filter) besides of its nutritional benefits (cereals, rich source of vitamin B<sub>6</sub>, phosphorus, iron, potassium, magnesium and thiamin) for the treatment of waste water.

# 2 Materials and methods

# 2.1. Reagents

All chemicals consumed in this work were of analytical grade. Stock solutions of fluoride and sulfide were prepared by dissolving a calculated amount of sodium fluoride (Sigma Aldrich) and sodium sulfide (Sigma Aldrich) in double distilled water to make the desired concentrations. For fluoride estimation 0.1 M HCl and NaOH solutions were used to adjust the *p*H of solution. For sulfide estimation 0.1 % of 1,10-phenanthroline and 0.01 M of Ferric chloride solutions were used. Acetate buffer was employed for controlling *p*H of solution.

# 2.2. Biosorbent

Rice bran (RB) was collected from a Rice Processing Mill, District Gujranwala, Punjab, Pakistan. Later, it was meshed to remove any foreign particles or impurities. One part of this raw rice bran (RRB) was used as such, whereas, remaining part was soaked in 0.1 M ZnCl<sub>2</sub> for 45 minutes and the resulting mixture was filtered, washed with distilled water, dried in oven for 24 h at 85°C. ZnCl<sub>2</sub> modified biosorbents were already reported to be used for removal of nitrate [28], sulphate [29], molybdate [30], thiocyanate ions [31, 32] and heavy metals [33, 34]. That's why this chemical treatment mode was selected. Later, this chemical treated rice bran (CRB) was ground into fine powder.

# **2.3.** Biosorption conditions optimization experiments

A series of experiments were performed with 50 mL solution of 10 ppm sodium fluoride and 50 mL solution of sodium sulfide (20 mg/L) solutions in batch mode, to optimize various parameters. Effects of various parameters on biosorption capacity was studied by varying biosorbent (RB and CRB) dose from 0.2-2.0 g, contact time 5-70 minutes, shaking speed 50-450 rpm, adsorbate concentration 2-14 ppm at room temperature (273 K). The effect of pH was studied in range of 3-9.



#### **Equipment and Apparatus**

The concentration of fluoride and sulfide in aqueous solutions was monitored spectrophotometrically using HITACHI U1800 spectrophotometer at 510 nm. FT-IR spectra were recorded in range of 4000 cm<sup>-1</sup> to 500 cm<sup>-1</sup> using Thermo Nicolet iS10 Spectrometer.

#### 2.5. Theoretical Analysis

It is important to determine the type of adsorption using various isotherm models. Various parameters of adsorption system are optimized by validating the experimental data. In this present work, experimental data were validated by using two well-known adsorption isotherm models. Adsorption isotherms determine the equilibrium associated between adsorbate and adsorbent.

Freundlich isotherm model is represented by logarithmic dependence at equilibrium of  $q_e$  (amount of adsorbate at equilibrium) on  $C_e$  (concentration of solution), equation (1):

$$\log q_e = \log K_F + \frac{1}{n} \log C_e \tag{1}$$

where " $K_F$ " is a adsorption capacity constant of the adsorbent (mg/g) and 1/n shows the adsorption intensity.

Langmuir isotherm model is represented by following equation (2):

$$1/q_e = 1/(bq_m c_e) + 1/q_m$$
 (2)

where " $C_e$ " denotes liquid phase equilibrium concentration (mg/L), " $q_m$ " is monolayer adsorption capacity (mg/g) and "b" is Langmuir constant (L/g), linked to the free adsorption energy.

#### 3. Results and discussions

Agricultural waste is a lignocellulosic material, used as adsorbent, because it is accessible in large amount and considered as a cheap and renewable material. Hence, a comparative study has been conducted to find the suitability and applicability of rice bran (raw and chemically treated) as ecofriendly adsorbent to remove fluoride and sulfide ions from aqueous solutions.

#### **3.1.** Characterization of (RRB) and (CRB) adsorbents

#### (a) FT-IR Analysis

A comparative study of FT-IR spectra of RRB and CRB has been done to identify various functional groups that act as potential adsorption sites. The recorded FT-IR spectra of raw (RRB), chemical treated rice bran (CRB) is shown in Figures1 and 2, respectively. The recorded peak shifts of RRB are shown in Table 1 [35]. The shifting of band 3287.63 cm<sup>-1</sup> (RRB) to 3273.52 cm<sup>-1</sup> (PU6) is representing hydroxyl (–OH) stretching. Peak 2918.88 cm<sup>-1</sup> (RRB) shifted to 2922.19 cm<sup>-1</sup> (PU6) and 2918.91 cm<sup>-1</sup> (PU7) is attributing the –CH stretching vibrations of its hydrogen, both asymmetric and symmetric. The peak 1709.27 cm<sup>-1</sup> (RRB) shifted to 1709.43 cm<sup>-1</sup> (PU7) is corresponding to C=O of aldehyde or carboxylic acid. The shift of peak 1651.63 cm<sup>-1</sup> (RRB) to 1650.53 cm<sup>-1</sup> (PU6) and 1648.35 cm<sup>-1</sup> (PU7) respectively is attributing the aromatic C=C, C=O and conjugated ketones or C=N amide stretching. Peak 1547.95 cm<sup>-1</sup> (RRB) is corresponding to symmetrical stretching of aromatic band C=O of aldehyde or carboxylic acid which disappeared in (PU6), (PU7). The methyl asymmetric C–H bands at peak 1242.15 cm<sup>-1</sup> (RRB) which also disappeared after fluoride and sulfide ions adsorbed. Peak 1015.99 cm<sup>-1</sup> (RRB) is shifted to 1031.72 cm<sup>-1</sup> (PU6) and 1029.27 cm<sup>-1</sup> (PU7) is ascribing the C–O stretching of alcohol, sulfoxides, carbohydrates or polysaccharide-like substances [36].





**Figure 1.** FT-IR spectra of raw rice bran (RRB) and post-adsorption of fluoride and sulfide ions on RRB



Figure 2. FTI-R spectra of chemical treated rice bran (CRB) and post-adsorption of fluoride and sulfide ions on CRB

**Table 1.** Comparative ft-ir analysis of raw rice bran (rrb) and Post-adsorption of fluoride and sulfide ions on rrb

FT-IR ANALYSIS, wavenumber (cm <sup>-1</sup> )					
<b>RRB</b> (B)*	PU6 (A)*	PU7 (A)*			
3387.63	3273.52	-			
2918.88	2922.19	2918.91			
2850.26	2833.12	2848.48			
1708.28	-	1709.43			
1651.63	1650.53	1648.35			
1547.95	-	-			
1242.15	-	-			
1015.99	1031.72	1029.27			

\*B (Before adsorption), \*A (After adsorption)



FT-IR ANALYSIS, wavenumber (cm <sup>-1</sup> )				
CRB (*	PU3 (A)*	PU5 (A)*		
2917.51	2918.14	2917.26		
2845.08	2845.88	2842.47		
1708.27	1708.97	1706.73		
-	1537.77	-		
-	-	1076.27		
1021.08	1018.75	1022.97		

<b>Table 2.</b> Comparative FT-IR analysis of chemical treated rice bran (crb) and
post-adsorption of fluoride and sulfide ions ON CRB

\*B (Before adsorption), \*A (After adsorption)

In case of CRB (Table 2), the shifting of band 2917.51 cm<sup>-1</sup> (CRB) to 2918.14 cm<sup>-1</sup> (PU3) and 2917.26 cm<sup>-1</sup> (PU5) is attributing the –CH asymmetric and symmetric stretching vibrations of methylene hydrogen. The peak 1708.27 cm<sup>-1</sup> (CRB) is shifted to 1708.97 cm<sup>-1</sup> (PU3) and 1706.73 cm<sup>-1</sup> (PU5) corresponding to C=O of aldehyde or carboxylic acid. Emergence of new park 1537.77 cm<sup>-1</sup> in PU3 is corresponding to symmetrical stretching of aromatic band C=O of aldehyde or carboxylic acid. The new absorption peak of 1076 cm<sup>-1</sup> in PU5 and peak 1021.08 cm<sup>-1</sup> (CRB) to 1018.75 cm<sup>-1</sup> (PU3) and 1022.97 cm<sup>-1</sup> (PU5) is ascribing the C–O stretching of alcohol, sulfoxides, carbohydrates or polysaccharide-like substances [37].

#### (b) SEM Analysis

The morphological changes occurred in rice bran sample before and after biosorption of sulfide and fluoride ions were monitored by scanning electron microscopy (SEM). SEM image of raw rice bran Fig. 3-a showed a gelatinous prickly pear morphology which transformed into mottled surface with micro fluffs after the adsorption of fluoride ions

Figure 3-b, whereas, the prickly pear texture of raw rice bran was filled after the adsorption of sulfide ion, giving an abraded morphology Figure 3-c.



**Figure 3.** (a) Raw rice bran before adsorption (b) Raw rice bran after the adsorption of  $F^-$  ions (c) Raw rice bran after the adsorption of  $S^{2-}$  ions



**Figure 4.** (a) Chemically treated rice bran (CRB) before adsorption (b) CRB after the adsorption of F<sup>-</sup> ions (c) CRB after the adsorption of S<sup>2-</sup> ions



SEM image of chemically treated rice bran showed a gelatinous morphology (Figure 4-a) which transformed into flaked surface after the adsorption fluoride ions (Figure 4-b), whereas, after the adsorption of sulfide ions mottled gel like morphology (Figure 4-c) [38].

# **3.2.** Study of Adsorption Process Characteristics

# (a) Effect of Adsorbent Dose

Particularly, the adsorbent dosage is important as it determines the degree of removal and it may be used to calculate the adsorbent cost per unit of aqueous solution to be treated. Generally, the percentage removal efficiency increases with an increase in adsorbent dosage, since more chelating sites are available. The same phenomenon was detected for the biosorption of fluoride and sulfide ions using RRB and CRB, as shown in Figure 5. In case of RRB, maximum biosorption 82.35% was observed for sulfide removal (SRRB), whereas, in case of CRB, maximum removal efficiency 78.61% was observed for fluoride removal (FCRB curve). It was evident that biosorption capacity enhanced by increased adsorbent dose because the surface area of adsorbent increases with availability of higher active adsorption sites.



**Figure 5.** Adsorbent dose effect on biosorptive elimination of fluoride and sulfide ions; 10 ppm fluoride ions, 20 mg/L sulfide ions, both solutions in 50 mL volume; RRB, CRB 0.1-2.0 g; time 5-70 min; shaking speed 50-450 rpm; *p*H 3-9; temperature 293K

#### (b) Effect of Contact Time

For the choice of an adequate adsorbent, contact time is important. The effect of contact time on fluoride and sulfide ions removal is shown in Figure 6. It is evident from the results that CRB was more effective in sulfide ion chelation with maximum biosorption efficiency 82.73% in 25 min (SCRB curve), whereas, CRB also showed maximum biosorption at 81.14% of fluoride ions in 35 min (FCRB curve). There was no noticeable change in biosorptive removal of anions after these optimized contact intervals, because under operating conditions, maximum adsorption took place due to the presence of higher number chelating site on biosorbent surface. Generally, the adsorbate molecules covered adsorbent surface progressively and after some time block the whole surface. When it occurs, the adsorbent unable to adsorb any more adsorbate molecules, which reflects that the adsorption capacity rises with time and reaches a constant value of time after which no considerable increase is removal efficiency observed.





Figure 6. Contact time effect on biosorptive elimination of fluoride and sulfide ions; 10 ppm fluoride ions, 20 mg/L sulfide ions, both solutions in 50 mL volume; RRB, CRB 0.1-2.0g; time 5-70 min; shaking speed 50-450 rpm; pH 3-9; temperature 293K

#### (c) Effect of Agitation Speed

It is shown Figure 7 that the maximum biosorption for CRB was 82.25% of sulfide ions at 200 rpm (SCRB curve), while CRB also showed 79.44% of fluoride ions biosorptive removal at 250 rpm (FCRB curve). In adsorption process, the agitation rate is an important parameter, as it effects the adsorbate distribution in the solution and the construction of the outer boundary film. The increase in agitation speed results in more turbulence and energy dissipation in the mixing zone, which decrease the boundary layer resistance and boundary film thickness. These facts increase the adsorption capacity at the early stages.



**Figure 7.** Agitation speed effect on biosorptive elimination of fluoride and sulfide ions; 10 ppm fluoride ions, 20 mg/L sulfide ions, both solutions in 50 mL volume; RRB, CRB 0.1-2.0 g; time 5-70 min; shaking speed 50-450 rpm; *p*H 3-9; temperature 293K

#### (d) Effect of pH

It is considered as the most important parameter that affects the adsorption process. This parameter affects the degree of ionization of the adsorbate in the solution, the surface charge of the adsorbent and the division of functional groups on the active sites of the adsorbent. The effect of pH on removal



efficiency is shown in Figure 8. CRB showed maximum removal efficiency 79% of fluoride ions (FCRB curve), at lower pH=5 which was possibly due to the rise in electrostatic attraction between positively charged functional groups and negatively charged anions molecule.



**Figure 8.** Waste water solution *p*H effect on biosorptive elimination of fluoride and sulfide ions; 10 ppm fluoride ions, 20 mg/L sulfide ions, both solutions in 50 mL volume; RRB, CRB 0.1-2.0 g; time 5-70 min; shaking speed 50-450 rpm; *p*H 3-9; temperature 293K

#### (e) Adsorption Isotherms

The adsorption isotherms are important from both practical and theoretical point of views. The Langmuir isotherm describes monolayer biosorptive removal of contaminants occurred by biosorbents having heterogeneously distributed chelating binding sites on their surfaces [39]. The correlation coefficients and values of parameters obtained from the adsorption of fluoride and sulfide ions on RRB and CRB are given in Table 3. It was revealed that data for adsorption of sulfide and fluoride onto raw and treated rice bran gave a good fit to the Langmuir adsorption isotherm thereby favoring the monolayer chelation mode of biosorption. Values of the regression coefficient also showed the better fit to monolayer adsorption. In case of zinc chloride treated rice bran the adsorption capacity increased for both the sulfide and fluoride. It means that increased roughness on the adsorbents surface provided more surface area for the adsorption of these adsorbates.

Adsorbent		Isotherm's parameters	
Pure	For fluoride or sulfide adsorption	Langmuir	Freundlich
RRB	FRRB	$R^2 = 0.994$ $q_m = 1.37 mg/g$ b = 0.03 L/g	$\begin{array}{l} R^2 = 0.990 \\ n = 0.93 \\ K_F = 0.082 \; (L/g)^{1\text{-}n} \end{array}$
	SRRB	$\begin{array}{l} R^2 = 0.99 \\ q_m = 3.79 \ mg/g \\ b = 0.099 \ L/g \end{array}$	$\begin{array}{l} R^2 = 0.99 \\ n = 1.10 \\ K_F = 0.33 \ (L/g)^{1\text{-}n} \end{array}$
CRB	FCRB	$\begin{array}{l} R^2 = 0.992 \\ q_m = 2.06 \ mg/g \\ b = 0.03 \ L/g \end{array}$	$\label{eq:rescaled} \begin{split} R^2 &= 0.987 \\ n &= 0.98 \\ K_F &= 0.050 \ (L/g)^{1\text{-n}} \end{split}$
	SCRB	$\begin{array}{l} R^2 = 0.993 \\ q_m = 7.08 \ mg/g \\ b = 0.013 \ L/g \end{array}$	$\label{eq:rescaled} \begin{split} R^2 &= 0.974 \\ n &= 1.03 \\ K_F &= 0.089 \; (L/g)^{1\text{-}n} \end{split}$

Table 3. A comparison of various parameters of different isotherm models

#### **4.**Conclusions

Adsorption of sulfide and fluoride onto rice bran during the present study revealed that better adsorption of fluoride occurred in slightly acidic medium whereas sulfide showed good adsorption in basic medium. Relatively lesser adsorbent dose was required after the treatment with zinc chloride in both the cases. Isothermal study revealed that adsorption of sulfide and fluoride onto chemically



modified rice bran followed Langmuir adsorption isotherm and does not fit the Freundlich isotherm well. From the present work, it was concluded that zinc chloride treated rice bran is the good alternate of raw rice bran for the efficient removal of sulfide and fluoride.

Acknowledgment. We are thankful to home institute for funding this project.

#### References

1.SUDHA, R., KALPANA, K., RAJACHANDRASEKAR, T., ARIVOLI, S., J.Chemistry, 4, 2007, p. 238

2.\*\*\*World Health Organization, Guidelines for Drinking-water Quality. First Addendum to the 4<sup>th</sup> edition, 2017

- 3. ARORA, H., CHATTOPADHYA, S., Ind.J.Environ.Health, 16, 1974, p. 140
- 4. FAN, C.S., LI, K.C., J.Cleaner Prod., 57, 2013, p. 335
- 5. PONSOT, I., FALCONE, R., BERNARDO, E., Ceramics Int., **39**, 2013, p. 6907
- 6. SHEN, F., CHEN, X., GAO, P., CHEN, G., Chem.Eng.Sci., 58, 2003, p. 987

7.DROUICHE, N., DJOUADI-BELKADA, F., OUSLIMANE, T., KEFAIFI, A., FATHI, J., AHMETOVIC, E., Desal.Water Treat., **51**, 2013, p. 5965

8. DAR, A., SHAFIQUE, U., ANWAR, J., ALI MUNAWAR, M., J.Sulfur Chem., 36, 2015, p. 187

9. EDWARDS, S., ALHARTHI, R., GHALY, A.E., Amer.J.Environ.Sci., 7, 2011, p. 295

10. VALEIKA, V., BELESKA, K., VALEIKIENĖ, V., Polish J.Environ.Stud., 15, 2006, p. 623

11.DROUICHE, N., AOUDJ, S., HECINI, M., GHAFFOUR, N., LOUNICI, H., MAMERI, N., J.Hazard.Mat., **169**, 2009, p. 65

12.RAO, N.M., BHASKARAN, C., J.Fluor.Chem., 41, 1988, p. 17

13. TAHAIKT, M., EL HABBANI, R., HADDOU, A.A., ACHARY, I., AMOR, Z., TAKY, M.,

ALAMI, A., BOUGHRIBA, A., HAFSI, M., ELMIDAOUI, A., Desalination, **212**, 2007, p. 46 14.AMOR, Z., BARIOU, B., MAMERI, N., TAKY, M., NICOLAS, S., ELMIDAOUI, A.,

Desalination, **133**, 2001, p. 215

15.AMOR, Z., MALKI, S., TAKY, M., BARIOU, B., MAMERI, N., ELMIDAOUI, A., Desalination, **120**, 1998, p. 263

- 16. TEMBHURKAR, A., DONGRE, S., J.Environ.Sci.Engin., 48, 2006, p. 151
- 17. FAN, X., PARKER, D.J., SMITH, M.D., Water Research, 37, 2003, p. 4929
- 18. MURUGANANTHAN, M., RAJU, G.B., PRABHAKAR, S., J.Hazard.Mat., 109, 2004, p. 37
- 19. MILLERO, F.J., LEFERRIERE, A., FERNANDEZ, M., HUBINGER, S., HERSHEY, J., Environ.Sci.Technol., 23, 1989, p. 209

20. FOSSO-KANKEU, E., MULABA-BAFUBIANDI, A., Phys.Chem.Earth, Parts A/B/C, 67, 2014, p. 242

- 21. JAIN, C.K., MALIK, D.S., YADAV, A.K., Environ. Processes, 3, 2016, p. 495
- 22. XU, X., GAO, B., JIN, B., YUE, Q., J.Mol.Liquids, 215, 2016, p. 565
- 23. HAGGERTY, G.M., BOWMAN, R.S., Environ.Sci.Technol., 28, 1994, p. 452

24. TIAN, Y., WU, M., LIU, R., WANG, D., LIN, X., LIU, W., MA, L., LI, Y., HUANG, Y., J.Hazard.Mat., 185, 2011, p. 93

25. KAMBLE, S.P., DIXIT, P., RAYALU, S.S., LABHSETWAR, N.K., Desalination, **249**, 2009, p. 687

26. DAIFULLAH, A., YAKOUT, S., ELREEFY, S., J.Hazard.Mat., 147, 2007, p. 633

27. SIVASANKAR, V., RAJKUMAR, S., MURUGESH, S., DARCHEN, A., J.Hazard.Mat., 225, 2012, p. 164

28. KHAN, M.A., AHN, Y.T., KUMAR, M., LEE, W., MIN, B., KIM, G., CHO, D.W., PARK, W.B., JEON, B.H., Separation Sci.Technol., **46**, 2011, p. 2575

- 29. NAMASIVAYAM, C., SANGEETHA, D., Desalination, 219, 2008, p. 1
- 30. NAMASIVAYAM, C., PRATHAP, K., Environ. Technol., 27, 2006, p. 923



31. OZKUTUK, E.B., ÖZALP, E., OZCAN, A.A., DILTEMIZ, S.E., Hacettepe J.Biol.Chem., 37, 2009, p. 207

- 32. NAMASIVAYAM, C., SANGEETHA, D., Chemosphere, 60, 2005, p. 1616
- 33. ANIRUDHAN, T.S., SUCHITHRA, P., Indian J.Chem.Technol., 17, 2010, p. 247

34. ITODO, A., ABDULRAHMAN, F., HASSAN, L., MAIGANDI, S., ITODO, H., New York Sci.J., **3**, 2010, p. 17

35. RAHMAN, N.N.N.A., SHAHADAT, M., WON, C.A., OMAR, F.M., RSC Advances, **4**, 2014, p. 58156

36. HENECZKOWSKI, M., KOPACZ, M., NOWAK, D., KUZNIAR, A., Acta Polon.Pharm. Drug Res., 58, 2001, p. 415

37. PAVIA, D.L, LAMPMAN, G.M, KRIZ, G.S., VYVYAN, J.R., Introduction to

Spectroscopy, 4th edition, Brooks/Cole, CENCAGE Learning, Belmont CA, 2009

38. FENG, W., KWON, S., BORGUET, E., VIDIC, R., Environ.Sci.Technol., 39, 2005, p. 9744

39. HO, Y.S., WANG, C.C., Process Biochem., 39, 2004, p. 761

#### List of Abbreviations:

- RRB = Raw rice bran
- CRB = Chemical treated rice bran
- FRRB = Fluoride biosorption on Raw rice bran
- FCRB = Fluoride biosorption on Chemical treated rice bran
- SRRB = Sulfide biosorption on Raw rice bran
- SCRB = Sulfide biosorption on Chemical treated rice bran



Grafical abstracts

#### Highlights

-Removal of fluoride and sulfide anionic species was investigated in this work by rice bran.

-Adsorption capacity of rice bran was enhanced by chemical treatment with ZnCl<sub>2</sub>.

-Structural changes were monitored by FT-IR and SEM micrographs.

-Fluoride and Sulfide were examined spectrophotometrically in samples and effective removal was observed.

Manuscript received: 12.12.2019